

# Manual for Marine Monitoring in the

# COMBINE

## Programme of HELCOM

### Part B

### General guidelines on **quality assurance** for monitoring in the Baltic Sea

Annex B-12

Technical note on the determination of heavy metals  
and persistent organic compounds in biota

Appendix 4

Technical note on the determination of  
trace metallic elements in biota



Last updated: 29.10.2012 (Annex number changed from Annex B 13 to Annex B 12)

# ANNEX B-12 TECHNICAL NOTE ON THE DETERMINATION OF HEAVY METALS AND PERSISTENT ORGANIC COMPOUNDS IN BIOTA

## ANNEX B-12, APPENDIX 4: TECHNICAL NOTE ON THE DETERMINATION OF TRACE METALLIC ELEMENTS IN BIOTA

1. Introduction.....	1
2. Working conditions .....	2
3. Pretreatment of laboratory ware and reagents; contamination control' .....	2
4. Sample pretreatment .....	3
5. Sample decomposition.....	3
6. Calibration .....	5
7. Determination .....	6
8. References.....	6
Attachment 1. Technical note on the determination of total mercury in marine biota by Cold Vapour Atomic Absorption Spectroscopy.....	9
1. Possibilities of using Cold Vapour Atomic Absorption Spectrometry in total mercury analysis .....	9
2. Sample Pretreatment .....	9
3. Control of Contamination and Analyte Losses .....	10
4. Reducing Reagents .....	11
5. Interferences .....	11
6. Internal (Routine) Quality Control.....	11
7. References .....	12

### 1. INTRODUCTION

Metallic elements appear in different marine biological matrices in trace concentrations, ranging from the mg/kg through the fĚg/kg to the ng/kg level. Stoeppler (1991) provided a comprehensive review of the most frequently used techniques for quantitative analysis of metallic trace elements, such as optical atomic absorption, fluorescence or emission spectrometry, anodic, cathodic or adsorptive stripping voltammetry,

isotope dilution mass spectrometry and total reflection X-ray fluorescence, respectively. In spite of the powerful instrumental techniques presently in use, various analytical error sources have to be taken into consideration that may significantly influence the accuracy of the analytical data.

---

## 2. WORKING CONDITIONS

For each step of the analytical procedure, contamination of the sample may occur from the environment (laboratory air dust particles and the analyst), from sample containers or packing materials, from instruments used during sample pre-treatment and sample preparation, and from the chemical reagents used for analysis.

The predominant purpose of the analytical clean laboratory is to eliminate contamination, which may be airborne or laboratory-induced, as far as possible and to control the total analytical blank. Contamination by particles from the laboratory air may be controlled by a high-efficiency particulate filter. (A clean room is designed to maintain air with 100 particles per ft<sup>3</sup> or 3.6.10<sup>3</sup> per m<sup>3</sup> of 0.5 fÊm particles (class 100 of U.S. Federal Standards 209), or better, preferably with a minimum of activity in the room.) U.S. Federal Standards 209 describes designs for complete laminar flow rooms, clean benches, and fume hoods, and contains information on design, testing, and maintenance of clean rooms, and should be considered an essential reference for those interested in a clean laboratory.

To control the analytical blank for analysis of metallic trace elements, one must not only maintain good laboratory air quality, but also select the appropriate composition and type of construction materials used to build the laboratory.

Principally, contaminants must be effectively removed at the source to minimize their uncontrolled distribution in the analytical clean laboratory. Accordingly, the laboratory's walls should be cleaned easily and therefore painted with special metal-free wipe-resistant paints. Surfaces of working areas should be protected with, for example, disposable plastic (polyethylene, PTFE) foils. The floors should, for example, be covered with adhesive plastic mats. Details of the design that are essential for obtaining a working laboratory with low trace element blanks are described by Moody (1982), Mitchell (1982a), Boutron (1990), and Schmidt and Gerwinski (1994).

---

## 3. PRETREATMENT OF LABORATORY WARE AND REAGENTS; CONTAMINATION CONTROL'

Chemically resistant materials, used in the production of high-quality laboratory ware appropriate for metallic trace element analysis, include low- and high-density polyethylene (LDPE and HDPE), polypropylene (PP), polytetrafluorethylene (PTFE), perfluoralkoxy (PFA), ethylenetetrafluorethylene (ETFE), tetrafluorethyleneper- fluorpropylene (FEP), borosilicate and quartz glass, respectively. With appropriate pretreatment and handling, these materials meet the requirements of purity necessary for the required analytical investigations. Cleaning procedures for plastic and glass laboratory ware were comprehensively dealt with by Moody and Lindstrom (1977), Tschopel et al. (1980), Kosta (1982) and Boutron (1990).

Generally, immersion in diluted (10-25 % v/v) high-purity nitric acid at room temperature for a period of one to three days, followed by repeated rinsing with high-purity water, is recommended.

Steaming in high-purity acids (predominantly nitric acid) is also very effective to remove impurities from container surfaces and condition them for subsequent analysis.

The materials mentioned above for the production of laboratory ware exhibit some adsorptive or exchange properties. Boundary-surface interactions can be important, particularly when very dilute analytical solutions are being handled, since uncontrollable losses through sorption of element ions can occur (Tschopel et al., 1980; Harms, 1985). Based on this information, it is imperative that volumetric flasks, reagent vessels, pipette tips, etc., for handling samples, sample solutions and low-level reference or analyte solutions must never be used for transferring or processing stock calibration solutions, analytes solutions or concentrated reagents. Considerable quantities of analytes may be adsorbed from such solutions by the respective container surfaces, residuals of which may be leached later when dilute sample or analyte solutions are handled.

The availability of high-purity reagents is a key condition for reliable investigations of metallic trace element concentrations. For many analytical problems, the level of a specific contaminant can adequately be controlled only by applying specific purification methods.

The first order of priority in regard to high-purity reagents is a sufficient supply of high-purity water. Ion-exchange units are universally accepted as an effective means of removing dissolved ionic species from water. Since high-purity water is frequently used in metallic trace element analysis, equipment for sustainable production of high-purity water by high-purity mixed-bed ion exchange resins should be available.

The next most important group of reagents are mineral acids. Contamination of the sample by residual concentrations of metallic trace elements in the acids used for dissolution or decomposition represents a major problem. Purification of the acids is essential to ensure acceptable blanks.

Isothermal (isopiestic) distillation can produce volatile acids (and ammonia) of medium concentration in high-purity form. For example, pure hydrochloric acid (and ammonia) can be generated by placing an open container of concentrated reagent-grade acid adjacent to a container of high-purity water, within a closed system (such as a desiccator) at room temperature. Acid vapours are continuously transferred into the water until equilibrium is obtained. Purification by sub-boiling distillation is based on motionless evaporation of the liquid by infrared heating at the surface to prevent violent boiling. Different purification systems are described in detail by Matthinson (1972), Kuehner et al. (1972), Dabeka et al. (1976), Tschopel et al. (1980), Mitchell (1982b), Moody and Beary (1982), Moody et al. (1989), and Paulsen et al. (1989). Acids of extremely high purity are produced by multiple batchwise distillation of reagent-grade acids in a silica apparatus, which is placed in a laminar-flow hood.

---

#### 4. SAMPLE PRETREATMENT

If the determinands are heterogeneously distributed in the sample material, it may be preferable to homogenize prior to taking subsamples for analysis. However, this procedural step is problematic, since uncontrollable contamination through the homogenizing tool may occur. Cryogenic homogenization at liquid nitrogen temperature and application of high-purity material such as quartz, PTFE, titanium or stainless steel for the construction of homogenizing devices may help to minimize contamination (Iyengar, 1976; Iyengar and Kasperek, 1977; Klussmann et al., 1985).

---

#### 5. SAMPLE DECOMPOSITION

For accurate direct measurements of metallic trace element contents in biological matrices, appropriate calibration (reference) standards are lacking in most instances. Therefore, multi-stage, easy to calibrate methods are still necessary, which include decomposition procedures and transformation of biological material into solution.

As a general rule wet sample is to be subject to decomposition procedures to avoid contamination or loss of determinands. A general sample decomposition procedure cannot be recommended due to the diverse composition of materials to be analysed, as well as to the different elements to be determined, and also because of the variety of possible analytical methods applied. However, the following minimum requirements should be met:

- complete destruction of all organic material of the sample,
- avoidance of determinand losses,
- avoidance of contamination.

Complete decomposition of the organic matrix is a prerequisite for a variety of the subsequently used instrumental determination techniques. Residual dissolved organic carbon from biological materials incompletely disintegrated after decomposition with nitric acid causes problems particularly in voltammetric and polarographic determinations. Both are sensitive to interference from chelating and electroactive organic components coexisting in incompletely decomposed samples during analysis (Pratt et al., 1988; Wurfels et al., 1987, 1989). Residual dissolved organic carbon compounds even of low molecular weight can change the equilibria in the spray chambers for sample introduction in atomic emission spectrometry (AES), optical emission spectrometry (OES), and atomic absorption spectrophotometry (AAS) by changing the viscosity of the sample solution. In such cases, comparison with pure aquatic calibration standard solutions can lead to erroneous results. In graphite furnace atomic absorption spectrophotometry (GFAAS), residual organic carbon may undergo complicated secondary reactions with the analyte prior to or during the atomization process. Such 'matrix interferences' alter the rate at which atoms enter the optical path relative to that obtained for an undisturbed element standard (Harms, 1985; and other references cited there).

The comparatively simple dry ashing method using a muffle furnace is problematic, since both uncontrollable losses of the determinands and contamination through contact with the furnace material may occur.

Both, application of a carefully developed and controlled temperature programme and modifying the matrix prior to the ashing procedure (addition of ashing aids agents) may be suitable to prevent losses of volatile elements (special analytical problems concerning mercury determination are described in Attachment 1). The use of special materials (quartz, titanium, stainless steel) for the construction of sample containers may be helpful to minimise contamination.

In the widely applied wet ashing procedure in open systems, the sample is treated with acids, mainly nitric, sulphuric and perchloric acids, in different ratios and under different conditions. Usually large quantities of reagents and voluminous apparatus with large surfaces are needed for complete destruction of the organic

material. Serious contamination problems (too high blank values) may arise, if insufficiently purified acids are used.

The rate of reaction and efficiency of acid decomposition increase substantially with elevated temperatures. Accordingly, closed-vessel techniques, using conventional heating or microwave energy, have an advantage over open systems. As a result of the closed systems with vessels manufactured of dense and very pure material (PTFE, PFA, quartz), loss of elements through volatilisation and contamination by desorption of impurities from the vessel surface are significantly reduced. In addition, since only small quantities of high-purity acid (usually nitric acid) need to be used, extremely low analytical blanks can be obtained.

Kingston and Jassie (1986, 1988) comprehensively considered the fundamental parameters governing closed vessel acid decomposition at elevated temperatures using a microwave radiation field. Microwave systems enable a very fast energy transfer to the sample and a very rapid build up of high internal vessel temperature and pressure, with the advantage of an enormous reduction in digestion time occurs. Furthermore, a reduction of acid volume (McCarthy and Ellis, 1991) and contamination reduction during the decomposition process were found (Dunemann, 1994; Sheppard et al., 1994).

The application of microwave energy must be carefully controlled to avoid explosions; a pressure-relief system is recommended for safe operation (Gilman and Grooms, 1988). At this stage of development, it can be concluded that advances in pressure and temperature feedback control features have contributed to the acceptance of microwave sample decomposition in analytical chemistry.

---

## 6. CALIBRATION

For calibration purposes, single element standard stock solutions at a concentration of 1000 mg/l, purchased from a qualified manufacturer, should be available. The actual concentration of the named element should be stated on the label together with the date of the preparation of the standard solution.

Fresh stock standard solutions should be compared with the old standard solutions. Traceability can be ensured by the use of CRM(s) or participation in intercomparison exercises (EURACHEM, 2003).

Single or mixed element working standard solutions for calibration purposes are prepared by dilution of the standard stock solutions using dilute acid, as required.

Both stock standard and working standard solutions are stored in polyethylene, borosilicate or quartz volumetric flasks. Working standard solutions at concentrations less than 100 µg/l should be freshly prepared for every batch of samples and kept no longer than two weeks.

The calibration procedure must meet some basic criteria in order to give the best estimate of the true (but unknown) element concentration of the sample analysed. These criteria are as follows:

- The amounts or concentrations of standards for the establishment of the calibration function must cover the range as related to practical conditions. The mean of the range should be roughly equal to the expected analyte concentration in the sample.

- The required analytical precision must be achievable and known throughout the entire range.
- The measured value (response) at the lower end of the range must be significantly different from the procedural analytical blank.
- The chemical and physical properties of the calibration standards must closely resemble those of the sample under investigation.
- The calibration standards must be processed through the entire analytical procedure in the same manner as the sample.
- The standard addition technique should be used only under very special circumstances (Cardone, 1986a, 1986b).

---

## 7. DETERMINATION

In an analytical series, especially with the number of samples >10, the control of calibration settings should be carried out with 2-3 calibration solution between environmental 10 samples. The analytical series should contain also a control sample of LRM or CRM.

---

## 8. REFERENCES

Boutron, C.F. 1990. A clean laboratory for ultralow concentration heavy metal analysis. *Fresenius Journal of Analytical Chemistry*, 337: 482-491.

Cardone, M. 1986a. New technique in chemical assay calculations. 1. A survey of calculational practices on a model problem. *Analytical Chemistry*, 58: 433-438.

Cardone, M. 1986b. New technique in chemical assay calculations. 2. Correct solution of the model problem and related concepts. *Analytical Chemistry*, 58: 438-445.

Dabeka, R.W., Mykytiuk, A., Berman, S.S., and Russell, D.S. 1976. Polypropylene for the sub-boiling distillation and storage of high-purity acids and water. *Analytical Chemistry*, 48: 1203-1207.

Dunemann, L. 1994. In *Probennahme und Aufschluss*, Ed. by M. Stoeppler. Springer, Heidelberg. p. 139.

Gilman, L., and Grooms, W. 1988. Safety concerns associated with wet ashing samples under pressure heated by microwave energy. *Analytical Chemistry*, 60: 1624-1625.

Harms, U. 1985. Possibilities of improving the determination of extremely low lead concentrations in marine fish by graphite furnace atomic absorption spectrometry. *Fresenius Journal of Analytical Chemistry*, 322: 53-56.

Iyengar, G. 1976. Homogenised sampling of bone and other biological materials. *Radiochemical and Radioanalytical Letters*, 24: 35-42.

- Iyengar, G., and Kasperek, K. 1977. Application of the Brittle Fracture Technique (BFT) to homogenise biological samples and some observations regarding the distribution behaviour of trace elements at different concentration levels in a biological matrix. *Journal of Radioanalytical Chemistry*, 39: 301-316.
- Kingston, H., and Jassie, L. 1986. Microwave energy for acid decomposition at elevated temperatures and pressures using biological and botanical samples. *Analytical Chemistry*, 58: 2534-2541.
- Kingston, H., and Jassie, L. (Eds.) 1988. Introduction to microwave sample preparation. ACS Professional Book, Washington, D.C.
- Klussmann, U., Strupp, D., and Ebing, W. 1985. Entwicklung einer Apparatur zur Homogenisierung von tiefgekühlten Pflanzenproben, *Fresenius Journal of Analytical Chemistry*, 322: 456-461.
- Kosta, L. 1982. Contamination as a limiting parameter in trace analysis. *Talanta*, 29: 985-992.
- Kuehner, E.C., Alvarez, R., Paulsen, P.J., and Murphy, T.J. 1972. Production and analysis of special high-purity acids purified by sub-boiling distillation. *Analytical Chemistry*, 44: 2050-2056.
- Matthinson, J.M. 1972. Preparation of hydrofluoric, hydrochloric, and nitric acids at ultralow lead levels. *Analytical Chemistry*, 44: 1715-1716.
- McCarthy, H.T., and Ellis, P.C. 1991. Comparison of microwave digestion with conventional wet ashing and dry ashing digestion for analysis of lead, cadmium, chromium, copper, and zinc in shellfish by flame atomic absorption spectroscopy. *Journal of the Association of Official Analytical Chemists*, 74: 566-569.
- Megginson, C., McKenzie, C., and Wells, D. 1994. Practical steps to improve the quality control of the chromatography for chlorobiphenyl and organochlorine pesticide analysis. *Marine Pollution Bulletin*, 29: 228-234.
- Mitchell, J.W. 1982a. State-of-the-art contamination control techniques for ultratrace elemental analysis. *Journal of Radioanalytical Chemistry*, 69: 47-105.
- Mitchell, J.W. 1982b. Purification of analytical reagents. *Talanta*, 29: 993-1002.
- Moody, J.R. 1982. NBS clean laboratories for trace element analysis. *Analytical Chemistry*, 54: 1358A-1376A.
- Moody, J.R., and Beary, E.S. 1982. Purified reagents for trace metal analyses. *Talanta*, 29: 1003-1010.
- Moody, J.R., and Lindstrom, R.M. 1977. Selection and cleaning of plastic containers for storage of trace element samples. *Analytical Chemistry*, 49: 2264-2267.
- Moody, J.R., Wissink, C.E., and Beary, E.S. 1989. Design principles for a large high-efficiency sub-boiling still. *Analytical Chemistry*, 61: 823-827.
- Paulsen, P.J., Beary, E.S., Bushee, D.S., and Moody, J.R. 1989. Analysis of ultrapure reagents from a large sub-boiling still made of teflon PFA. *Analytical Chemistry*, 61: 827-830.

Pratt, K., Kingston, H., MacCrehan, W., Koch, W. 1988. Voltammetric and liquid chromatographic identification of organic products of microwave-assisted wet ashing of biological samples. *Analytical Chemistry*, 60: 2024-2027.

Schmidt, D., and Gerwinski, W. 1994. Design principles of clean laboratories for trace metal analysis. In ICES/HELCOM Workshop on Quality Assurance of Chemical Analytical Procedures for the Baltic Monitoring Programme. Ed. by G. Topping, and U. Harms. *Baltic Sea Environment Proceedings No. 58*: 111-117.

Sheppard, B.S., Heitkemper, D.T., and Gaston, C.M. 1994. Microwave digestion for the determination of arsenic, cadmium and lead in seafood products by inductively coupled plasma atomic emission and mass spectrometry. *Analyst*, 119: 1683-1686.

Stoeppler, M. 1991. Analytical chemistry of metals and metal compounds. In *Metals and Their Compounds in the Environment*. Ed. by E. Merian. VCH Weinheim. pp. 105-206.

Tschopel, P., Kotz, L., Schulz, W., Veber, M., and Tolg, G. 1980. Zur Ursache und Vermeidung systematischer Fehler bei Elementbestimmungen in wasrigen Losungen im ng/ml- und pg/ml-Bereich. *Fresenius Journal of Analytical Chemistry*, 203: 1-14.

Wurfels, M., Jackwerth, E., and Stoeppler, M. 1987. Sample pretreatment studies with biological and environmental materials. The problem of disturbances of inverse voltammetric trace analysis after pressure decomposition of biological samples. *Fresenius Journal of Analytical Chemistry*, 329: 459.

Wurfels, M., Jackwerth, E., and Stoeppler, M. 1989. Residues from biological materials after pressure decomposition with nitric acid. *Analytica Chimica Acta*, 226: 1-41 (Part 1-3).

## ATTACHMENT 1. TECHNICAL NOTE ON THE DETERMINATION OF TOTAL MERCURY IN MARINE BIOTA BY COLD VAPOUR ATOMIC ABSORPTION SPECTROSCOPY

### 1. POSSIBILITIES OF USING COLD VAPOUR ATOMIC ABSORPTION SPECTROMETRY IN TOTAL MERCURY ANALYSIS

The most widely used method for the determination of total mercury in biological tissues is cold vapour atomic absorption spectrometry (CV-AAS), based on a technique elaborated in detail by Hatch and Ott (1968). In this method, (divalent) ionic mercury is reduced to its metallic form ( $\text{Hg}^0$ ) in acidic solution using a powerful reducing agent. Subsequently, the elemental mercury is volatilized (purged) by a carrier gas and transported into an absorption cell, where the 253.65 nm wavelength absorbance of mercury atoms is measured.

CV-AAS analysis can be performed manually using batch CV-AAS or automatically using flow injection (FI) techniques. FI is a very efficient approach for introducing and processing liquid samples in atomic absorption spectrometry. The FI technique, combined with a built-in atomic absorption spectrometer optimised for mercury determination, reduces sample and reagent consumption, has a higher tolerance of interferences, lower determination limits and improved precision compared with conventional cold vapour techniques.

The efficiency of various flow injection mercury systems has been reported by several groups (Tsalev *et al.*, 1992a, 1992b; Welz *et al.*, 1992; Guo and Baasner, 1993; Hanna and McIntosh, 1995; Kingston and McIntosh, 1995; Lippo *et al.*, 1997).

Better sensitivities of both conventional CV-AAS and FI-CV-AAS can be obtained by collecting mercury vapour released from the sample solution on a gold adsorber (Welz and Melcher, 1984). This so-called amalgamation technique eliminates kinetic interferences due to a different vaporization rate or a different distribution function of the elemental mercury between the liquid and the gaseous phases. The amalgamation ability of the gold adsorber must be carefully and regularly checked. Volatile compounds (in particular sulfur-containing compounds) evaporating together with the elemental mercury from the sample solution may deactivate the adsorber surface. This means an increased risk of underestimation, as unknown quantities of mercury are not collected by the adsorber.

### 2. SAMPLE PRETREATMENT

It is generally agreed that oxidative conversion of all forms of mercury in the sample to ionic  $\text{Hg(II)}$  is necessary prior to reduction to elemental Hg and its subsequent measurement by CV-AAS. Therefore, the initial procedural step in mercury analysis is a sample pretreatment, which is aimed at liberating the analyte element from its chemical bonding to the organic matrix and thus transforming all of the analyte species into a well-defined oxidation state. For this purpose, a wide variety of combinations of strong acids (HCl,  $\text{H}_2\text{SO}_4$ ,  $\text{HNO}_3$ ) and oxidants ( $\text{H}_2\text{O}_2$ ,  $\text{KMnO}_4$ ,  $\text{K}_2\text{Cr}_2\text{O}_7$ ,  $\text{K}_2\text{S}_2\text{O}_8$ ) have been tested and recommended (Kaiser *et al.*, 1978; Harms, 1988; Vermeiret *et al.*, 1989; Ping and Dasgupta, 1989; Baxter and Frech, 1990; Landi *et al.*, 1990; Navarro *et al.*, 1992; Lippo *et al.*, 1997).

A suitable sample pretreatment, which implies the complete transformation of all organomercury species into inorganic mercury ions, requires the following:

- oxidation mixtures with a high oxidation potential;
- rapid oxidation (usually promoted by high reaction temperatures), preferably in closed systems;
- compatibility with CV-AAS techniques;
- stability of sample solutions during storage (at least short term);
- no formation of solid reaction products.

On-line sample pretreatment is of particular interest in total mercury determinations because it allows reduction of the well-known problems associated with the inherent risk of contamination, and volatilization and adsorption losses. At present, suitable procedures for on-line pretreatment of solid biological samples are lacking. However, several authors (Tsalev *et al.*, 1992a 1992b; Welz *et al.*, 1992; Guo and Baasner, 1993) have demonstrated that microwave digestion coupled with FI-CV-AAS can successfully be applied to the analysis of liquid samples.

### 3. CONTROL OF CONTAMINATION AND ANALYTE LOSSES

Major difficulties arise due to the mobility and reactivity of mercury and its compounds, respectively, during sample preparation, sample pretreatment, and analysis. Therefore, the stability of samples and standard solutions is of prime importance, and it is advisable to test the stability of typical standard and sample solutions under typical laboratory conditions.

Mercury can disappear from solution due to several mechanisms, including volatilization of mercury compounds, reduction of such compounds followed by volatilization of elemental (metallic) mercury, adsorption on container walls, adsorption onto colloids or particles, incorporation into stable chemical complexes, or incorporation, upon reduction, into stable amalgams.

Thermodynamic considerations of Toribara *et al.* (1970) showed that loss of mercury from a solution containing the element in the monovalent form may occur readily through disproportion and subsequent loss of metallic mercury. Because of the high oxidation potential of the mercury(II)-mercury(I) system, almost any reducing substance could convert some divalent mercury ions into monovalent mercury ions, which then spontaneously disproportionate into mercury(II) and mercury(0). The latter escape as metallic vapour from the solution into the gas phase. Because of the almost impossibility of preventing the introduction of small amounts of reducing substances by reagents or solvents, the more dilute mercury(II) solutions would be less stable and lose mercury more readily. The only practical method for stabilizing such solutions is to add a small excess of an oxidising substance (such as permanganate), which has a higher oxidation potential than the mercury(II)-mercury(I) system.

Similarly, Feldman (1974) concluded from his experiments that solutions with 0.1 µg divalent Hg dm<sup>-3</sup> in distilled water could be stored in glass vials for as long as five months without deteriorating if the solutions contained 5 % (v/v) HNO<sub>3</sub> and 0.01 % Cr<sub>2</sub>O<sub>7</sub><sup>2-</sup>. Storage of such solutions was safe in polyethylene vials for at

least 10 days if the solutions contained 5 % (v/v) HNO<sub>3</sub> and 0.05 % Cr<sub>2</sub>O<sub>7</sub><sup>2-</sup>. The efficiency of this mixture was probably due to its ability to prevent the hydrolysis of dissolved mercury and prevent its reduction to valencies lower than +2.

#### 4. REDUCING REAGENTS

Tin(II) chloride and sodium tetrahydroborate are predominantly used as reducing reagents for the determination of total mercury by CV-AAS. Sodium tetrahydroborate has been found advantageous for several applications owing to its higher reducing power and faster reaction (Toffaletti and Savory, 1975). In addition, this reductant has been successfully used even in the presence of interfering agents such as iodide and selenium (Kaiser *et al.*, 1978). However, potential interferences can occur from metal ions (e.g., Ag(I), Cu(II), Ni(II)), which are themselves reduced to the metallic state and so may occlude mercury through amalgamation.

Welz and Melcher (1984) showed that sodium tetrahydroborate could more readily attack those organic mercury compounds which were not reduced to metallic mercury by tin(II) chloride. However, they stated that sodium tetrahydroborate could not be recommended as the reducing reagent for the amalgamation technique. They found that, due to the rather violent reaction with sodium tetrahydroborate, fine droplets of the sample solution were carried by the gas stream and contaminated or deactivated the adsorber surface. Further, they considered even more important the fact that not only mercury but all gaseous hydride-forming elements (e.g., arsenic, antimony, selenium) were volatilized when sodium tetrahydroborate was used as reductant. These hydrides reacted with the adsorber material and deactivated its surface, thus no longer permitting a sensitive and reproducible determination of mercury.

#### 5. INTERFERENCES

Interferences by volatile nitrogen oxides in the determination of mercury by FI-CV-AAS were studied by Rokkjaer *et al.* (1993). The main symptom of the interference effects was a suppression, broadening or even splitting of the mercury signal. The authors postulated that volatile nitrogen oxides formed as reaction products of nitric acid during sample decomposition scavenged the reducing agent and concomitantly inhibited the reduction of mercury(II). The rate of the reaction of nitrogen oxides with the reducing agent was considered to be so fast that it was consumed before the reduction of mercury was complete. Rokkjaer *et al.* (1993) demonstrated that the interference could easily be remedied by purging the sample solution with an inert gas prior to the introduction of the reducing agent. Lippo *et al.* (1997) concluded from their experiments that nitrogen mono- and dioxide, having molecular absorption bands at 253.63 nm and 253.85 nm, respectively, might cause unspecific absorption at the specific mercury wavelength of 253.65 nm, leading to enhanced and broadened mercury signals if not properly compensated for by adequate instrumental background correction.

#### 6. INTERNAL (ROUTINE) QUALITY CONTROL

In order to demonstrate that the analytical method applied is fit for the purpose of the investigations to be carried out, control materials should be regularly analysed alongside the test materials (cf. Chapter B.5 of the Manual).

The control materials - preferably certified reference materials (CRM) - should be typical of the test materials under investigation in terms of chemical composition, physical properties and analyte concentration. Fitness for purpose is achieved if the results obtained from the analysis of the control materials are within the defined limits of permissible tolerances in analytical error (see Chapters B.3.5, B.4.2.5 and B.4.2.5.2b of the Manual).

---

## 7. REFERENCES

- Baxter, D., and Frech, W. 1990. Critical comparison of two standard digestion procedures for the determination of total mercury in natural water samples by cold vapour atomic absorption spectrometry. *Analytica Chimica Acta*, 236: 377-384.
- EURACHEM 2003. EURACHEM/CITAC Guide "Traceability in Chemical Measurement" – A guide to achieving comparable results in chemical measurement. S.L.R. Ellison, B. King, M. Rösslein, M. Salit, A. Williams (Eds.), 43 pp.
- Feldman, C. 1974. Preservation of dilute mercury solutions. *Analytical Chemistry*, 46: 99-102.
- Guo, T., and Baasner, J. 1993. Determination of mercury in urine by flow-injection cold vapour atomic absorption spectrometry. *Analytica Chimica Acta*, 278: 189-196.
- Hanna, C., and McIntosh, S. 1995. Determination of total Hg in environmental samples with on-line microwave digestion coupled to a flow injection mercury system (FIMS). *Atomic Spectroscopy*, 16: 106-114.
- Harms, U. 1988: Analytical procedures for mercury, cadmium and lead in biological material. *Baltic Sea Environment Proceedings*, No. 27C, Part CI, pp. 36-62. Helsinki Commission.
- Hatch, W., and Ott, W. 1968. Determination of sub-microgram quantities of mercury by atomic absorption spectrometry. *Analytical Chemistry*, 40: 2085-2087.
- Kaiser, G., Götz, D., Tölg, G., Knapp, G., Maichin, B., and Spitzky, H. 1978. Untersuchung von systematischen Fehlern bei der Bestimmung von Hg-Gesamtgehalten im Bereich 10 -5 % in anorganischen und organischen Matrices mit zwei unabhängigen Verbundverfahren. *Fresenius Z. Analytical Chemistry*, 291: 278-291.
- Kingston, K., and McIntosh, S. 1995. Determination of mercury in geological samples by flow injection AAS. *Atomic Spectroscopy*, 16: 115-117.
- Landi, S., Fagioli, F., Locatelli, C., and Vecchiotti, R. 1990. Digestion method for the determination of mercury in vegetable matrices by cold vapour atomic absorption spectrometry. *Analyst*, 115: 173-177.
- Lippo, H., Jauhiainen, T., and Perämäki, P. 1997. Comparison of digestion methods for the determination of total mercury in environmental samples by flow injection CV-AAS. *Atomic Spectroscopy*, 18: 102-108.
- Navarro, M., Lopez, M., Lopez, H., and Sanchez, M. 1992. Microwave dissolution for the determination of mercury in fish by cold vapour atomic absorption spectrometry. *Analytica Chimica Acta*, 257: 155-158.

- Ping, L., and Dasgupta, P. 1989. Determination of total mercury in water and urine by gold film sensor following Fenton's Reagent digestion. *Analytical Chemistry*, 61: 1230-1235. Rokkjaer, J., Hoyer, B., and Jensen, N. 1993. Interference by volatile nitrogen oxides in the determination of mercury by flow injection cold vapour atomic absorption spectrometry. *Talanta*, 40: 729-735.
- Toffaletti, J., and Savory, J. 1975. Use of sodium borohydride for determination of total mercury in urine by atomic absorption spectrometry. *Analytical Chemistry*, 47: 2091-2095.
- Toribara, T., Shields, C., and Koval, L. 1970. Behaviour of dilute solutions of mercury. *Talanta*, 17: 1025-1028. Tsalev, D., Sperling, M., and Welz, B. 1992a: On-line microwave sample pre-treatment for hydride generation and cold vapour atomic absorption spectrometry. Part 1. The manifold. *Analyst*, 117: 1729-1733.
- Tsalev, D., Sperling, M., and Welz, B. 1992b. On-line microwave sample pre-treatment for hydride generation and cold vapour atomic absorption spectrometry. Part 2. Chemistry and application. *Analyst*, 117: 1735-1739.
- Vermeir, G., Vandecasteele, C., and Dams, R. 1989. Microwave dissolution for the determination of mercury in biological samples. *Analytica Chimica Acta*, 220: 257-261.
- Welz, B., and Melcher, M. 1984. Picotrace determination of mercury using the amalgamation technique. *Atomic Spectroscopy*, 5: 37-42.
- Welz, B., Tsalev, D., and Sperling, M. 1992. On-Line microwave sample pre-treatment for the determination of mercury in water and urine by flow-injection cold-vapour atomic absorption spectrometry. *Analytica Chimica Acta*, 261: 91-103.